

# Ozonation as a pre-treatment for anaerobic digestion of waste activated sludge: Effect of the ozone doses

G. Silvestre\*, M.B Ruiz\*, M. Fiter\*, C. Ferrer\*\*, J.G Berlanga\*\*, S. Alonso\*\*, A.Canut\*

\* AINIA Technology Centre, Valencia Technology Park, Benjamin Franklin 5-11, 46980 Paterna, Valencia, Spain (E-mail: {*acanut*}{*mfit*}{*bruiz*}{*gsilvestre*}@*ainia.es*)

\*\* Facsa, C/Mayor, 82-84. E-12001 Castelló, Spain.  
(E-mail: {*cferrer*}{*jgberlanga*}{*salonso*}@*facsa.com*)

## Abstract

The effect of different ozone doses was assessed in terms of biogas production, anaerobic digestion (AD) kinetics and concentration of amino acids (AA) and long chain fatty acids (LCFA) in the waste activated sludge (WAS). Four different ozone ( $O_3$ ) doses were used: 0.043  $gO_3/g_{TSS}$ , 0.063  $gO_3/g_{TSS}$ , 0.080  $gO_3/g_{TSS}$  and 0.100  $gO_3/g_{TSS}$ . The lower doses resulted in biogas production increases and faster kinetics in the AD of SS, while the higher doses led to poorer process performance. Ozonation of WAS resulted in a non-linear reduction of the AA concentration. The concentration of LCFA was also reduced with linear correlation with the ozone dose. The reaction products of LCFA (aldehydes) might be the cause of inhibition observed in the anaerobic digestion of WAS treated with higher ozone doses.

## Keywords

ozone; pre-treatment; WAS anaerobic digestion; amino acids; long chain fatty acids; aldehydes

## INTRODUCTION

In recent years, optimisation of the anaerobic digestion (AD) of sewage sludge (SS) has aroused great interest due to the use of biogas as a renewable energy source for increasing the energy self-sustainability of waste water treatment plants (WWTPs). One of the strategies for AD optimisation is the pre-treatment of waste activated sludge (WAS) to increase biogas production, reduce volatile solids, improve SS dewatering and reduce pathogen content. Biological, mechanical, thermal and chemical sludge disintegration has been investigated (Carrère *et al.*, 2010).

Ozonation is the chemical method most widely used to solubilise WAS. While ozonation applied to reduce excess sludge has been widely investigated, only few studies are focused in the ozonation of WAS with the aim to increase the biogas production during the AD process. The effect of ozonation on WAS AD depends on the  $O_3$  doses applied. Some controversy about the optimal  $O_3$  dose is seen in the bibliography. Yeom *et al.* (2003) found that the optimal dose to enhance biogas production was 0.2  $gO_3/g_{TSS}$  while Bougrier *et al.* (2006) suggested that mineralization of dissolved organic matter occurred at 0.15  $gO_3/g_{TS}$ .

The aim of this work is to increase knowledge on the effect of the  $O_3$  dose on the subsequent WAS AD. Biogas yield and kinetics were determined in batch AD of WAS pre-treated with four different  $O_3$  doses and compared with SS without pre-treatment. The best and worse doses in terms of biogas yield and kinetics were selected to analyze and compare the effect of  $O_3$  doses on the proteins and lipids present in WAS.

## MATERIALS & METHODS

### Substrates

Samples of WAS, primary sludge (PS) and inoculum for the experimental AD tests

were taken from the flotation tank, the primary settler and the anaerobic digester of Castelló de la plana WWTP respectively.

### **Ozonation**

The ozonation unit consists of a corona discharge type ozone generator with a production capacity of 2gO<sub>3</sub>/h from oxygen. The inlet gas preparation system consists of an oxygen concentrator TOPAZ plus model provided by AIRSEP. O<sub>3</sub> concentration in the O<sub>3</sub> enriched gas produced by this system is measured by means of an O<sub>3</sub> analyser GM-PRO (Anseros®). The contact reactor consists of a 10 L transparent PVC reactor. WAS was pre-treated applying 4 different doses (0.043 gO<sub>3</sub>/g<sub>TSS</sub>, 0.063 gO<sub>3</sub>/g<sub>TSS</sub>, 0.080 gO<sub>3</sub>/g<sub>TSS</sub> and 0.100 gO<sub>3</sub>/g<sub>TSS</sub>). Before the ozonation pre-treatment, the WAS samples were bubbled with N<sub>2</sub> until the oxygen content was negligible.

### **Batch anaerobic digestion test**

Two sets of batch AD tests were run using as a substrate 1) WAS (1 sample without pre-treatment and 4 ozonated samples) and 2) mixed sludge (60WAS:40PS vv) with the same pre-treatment scheme applied to the SS fraction. Experimental tests procedure was based on Verein Deutscher Ingenieure (VDI) Standard 4630 “Fermentation of organic materials”. Glass bottles with a capacity of 2L were filled with inoculum and the substrate with an inoculum to substrate ratio (ISR) of 1 in terms of volatile solids. The vials were bubbled with N<sub>2</sub> and placed inside an oven at 33°C. Biogas production was measured by Ritter Milligascounters® MGC-1 flow meter. Tests were run in triplicates.

### **Analytical methods**

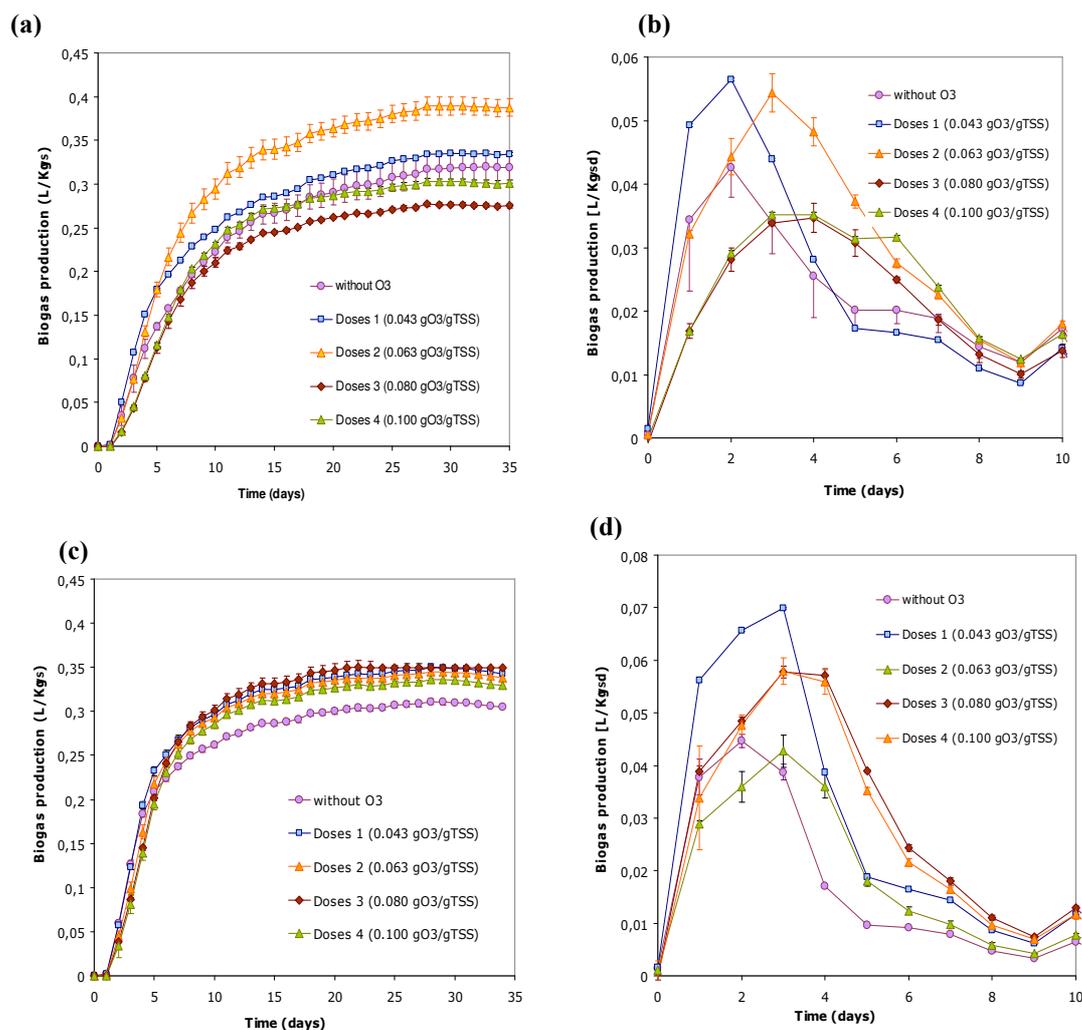
Total solids (TS), volatile solids (VS), total and volatile suspended solids (TSS and VSS) were determined according to Standard Methods (APHA, AWWA, WEF, 1998). AA were analyzed by HPLC-PDA and LCFA on the fat fraction were determined by GC-FID.

## **RESULTS**

### **Effect of ozone on biogas potential**

*Secondary sludge.* Cumulative biogas production curves and daily biogas production of the 10 first days of batch AD test of the WAS are shown in figures 1a and 1b respectively. Dose 1 (0.043 gO<sub>3</sub>/g<sub>TSS</sub>) shows the faster kinetics, reaching the maximum daily biogas production by the second day of experiment (figure 1b). The biogas potential was increased by 5% with respect to the WAS without ozone pre-treatment. Dose 2 (0.063 gO<sub>3</sub>/g<sub>TSS</sub>) shows similar kinetics as the non pre-treated WAS sample, but the biogas potential is 21% higher. Doses 3 and 4 (0.080 & 0.100 gO<sub>3</sub>/g<sub>TSS</sub>) showed worse –and similar– initial kinetics and less biogas potential than the non pre-treated SS, but dose 4 had higher biogas potential than dose 3. The daily biogas production of SS treated with dose 4 increased by the sixth day of the experiment suggesting that some extra biodegradable organic matter was available. Biogas composition (CH<sub>4</sub>, CO<sub>2</sub>, H<sub>2</sub>S, and H<sub>2</sub>) was similar in all samples (data not shown).

*Mixed sludge.* Cumulative biogas production curves and daily biogas production of the mixed sludge are shown in the figures 1c and 1d. In this case the mixtures of PS with the pre-treated WAS showed a similar increase in the biogas potential with respect to the mixture with WAS without pre-treatment, regardless of the O<sub>3</sub> dose applied. This fact suggested that the lower biogas potential of the WAS pre-treated with the higher doses 3 and 4 (figures 1a and 1b) is due to an inhibitory effect and not to a mineralization of the organic matter.



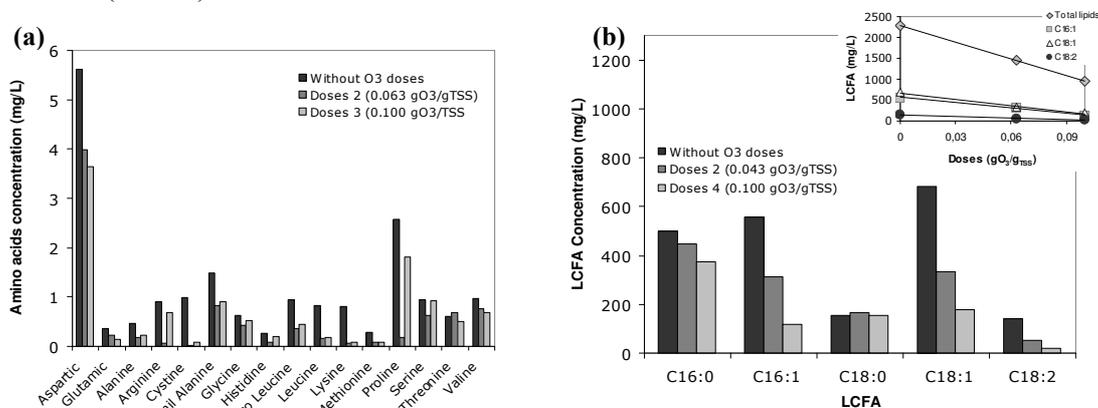
**Figure 1.** (a) Cumulative biogas production of WAS; (b) Daily biogas production of WAS; (c) Cumulative biogas production of mixed sludge; (d) Daily biogas production of mixed sludge.

### Effect of ozone on the amino acids and long chain fatty acids profiles

**Amino acids.** Figures 2a and 2b show respectively the total AA concentration and the LCFA concentration in the WAS non pre-treated and ozonated with dose 2 (0.063 gO<sub>3</sub>/gTSS) and 4 (0.100 gO<sub>3</sub>/gTSS). The total and soluble AA concentrations were 19 mg/L and 3 mg/L (no pre-treatment), 10 mg/L and 1 mg/L (dose 2, 0.063 gO<sub>3</sub>/gTSS) and 12 mg/L and 1 mg/L (dose 4, 0.100 gO<sub>3</sub>/gTSS). These results show that O<sub>3</sub> has an effect on the total and soluble AA. O<sub>3</sub> might introduce changes in their secondary and tertiary structure and it is possible that these changes are connected with the partial oxidation of the aromatic monomeric units of the proteins (Cataldo, 2003). The results suggested that the O<sub>3</sub> doses have not a linear relation with the AA profile.

**Long chain fatty acids.** The total lipids concentration in the WAS were 2275 mg/L (no pre-treatment), 1450 mg/L (dose 2, 0.063 gO<sub>3</sub>/gTSS) and 959 mg/L (dose 4, 0.100 gO<sub>3</sub>/gTSS). O<sub>3</sub> reacts with the unsaturated LCFA and the results suggested that a linear correlation exists between the LCFA concentration and the O<sub>3</sub> doses in the range studied in this work. In the presence of water, the products obtained from the LCFA oxidation with O<sub>3</sub> are aldehydes and hydroxyhydroperoxide (Pryor *et al.*, 1995). Aldehydes are inhibitors for the AD (Chen *et al.*, 2008). These results could explain the results observed in the batch AD experiments; at high O<sub>3</sub> doses (dose 3 and 4) the inhibitor

concentration would be higher in the AD of SS only, thus causing a decrease on the biogas production, while in the AD of mixed sludge, the inhibitor would be diluted and therefore no such effects are observed. Moreover, if the O<sub>3</sub> dose is high enough, the aldehydes might be further oxidized and therefore their inhibitory effect would be reduced (dose 4).



**Figure 2.** (a) AA concentration in the WAS with and without pre-treatment (dose 2 and 4); (b) LCFA concentration in the WAS with and without pre-treatment (dose 2 and 4).

## CONCLUSIONS

The effect of ozonation on the biogas production and anaerobic digestion kinetics of waste activated sludge is different according to the ozone dose applied. The lower doses applied in this work led to improvements of the biogas production and faster kinetics, but poorer performance was observed with higher ozone doses. This effect is not observed in the anaerobic digestion of mixed sludge with ozonated waste activated sludge.

Ozone reduces the concentration of total AA and LCFA in the waste activated sludge. The LCFA oxidation produces aldehydes, which are inhibitory of the AD. This could explain the results of AD tests. Further research will be done to confirm this hypothesis.

## ACKNOWLEDGEMENTS

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## REFERENCES

- Bougrier, C., Albasi, C., Delgenès, J.P., Carrère, H., 2006. Effect of ultrasonic, thermal and ozone pre-treatments on waste activated sludge solubilisation and anaerobic biodegradability. *Chemical Engineering and Processing* **45** (8), 711–718.
- Carrère H., Dumas, C., Battimelli, D.J., Batstone, J.P., Delgenès, J.P., Steyer, I., Ferrer 2010. *Journal of Hazardous Materials* **183**, 1-15.
- Cataldo, F., 2003. On the action of ozone on proteins. *Polymer Degradation and Stability* **82**, 105-114
- Chen Y., Cheng J.J., Creamer K.S., 2008. Inhibition of anaerobic digestion process: a review. *Bioresource Technology* **99**, 4044-4064.
- Pryor, W.A., Squadrito G.L., Friedman M., 1995. The cascade mechanism to explain ozone toxicity : the role of lipid ozonation products. *Free Radical Biology & Medicine*, **19**, 935-941.
- Yeom, I.T., Lee, K.R., Lee, Y.H., Ahn, K.H., Lee, S.H., (2002). Effects of ozone treatment on the biodegradability of sludge from municipal wastewater treatment plants. *Water Science and Technology* **46** (4–5), 421–425.